53

55

60

65

70

75

80

1

0 **Response of pollinators to the tradeoff between resource** acquisition and predator avoidance

# Ana L. Llandres, Eva De Mas and Miguel A. Rodríguez-Gironés

A. L. Llandres (allandres@eeza.csic.es), E. De Mas and M. A. Rodríguez-Gironés, Dept of Functional and Evolutionary Ecology, Estación Experimental de Zonas Áridas (CSIC), Carretera de Sacramento, s/n, ES-04120, La Cañada de San Urbano, Almería, Spain.

10

5

Although the behaviour of animals facing the conflicting demands of increasing foraging success and decreasing predation risk has been studied in many taxa, the response of pollinators to variations in both factors has only been studied in isolation. We compared visit rates of two pollinator species, hoverflies and honeybees, to 40 Chrysanthemum segetum patches in which we manipulated predation risk (patches with and without crab spiders) and nectar availability (rich and poor patches) using a full factorial design. Pollinators responded differently to the tradeoff between maximising intake rate and minimising predation risk: honeybees preferred rich safe patches and avoided poor risky patches while the number of hoverflies was highest at poor risky patches. Because honeybees were more susceptible to predation than hoverflies, our results suggest that, in the presence of competition for resources, less susceptible pollinators concentrate their foraging effort on riskier resources, where competition is less severe. Crab spiders had a negative effect on the rate at which inflorescences were visited by honeybees. This effect was mediated through changes in the foraging strategy of honeybees, and could, in principle, be reversed by increasing nectar productivity of inflorescences. Our study shows that both pollinator species responded simultaneously and differently to variations in food reward and predation risk, and highlights the importance of studying the foraging strategies of pollinators in order to fully understand how plant-pollinator interactions are established.

- 25 Because of their role as pollen vectors, pollinators have the potential to influence the structure of plant communities. At the same time, however, plant-pollinator interactions represent just a step in more complex food webs and cannot be studied in isolation (Jordano 1987, Bascompte
- 30 et al. 2003, Ings et al. 2009a). A proper understanding of plant-pollinator interactions therefore requires elucidating the factors determining the foraging choices of pollinators (Waser and Price 1998, Rodríguez-Gironés and Santamaría 2010). While the factors affecting foraging choices of
- 35 pollinators can be studied both in the field and the laboratory (Pleasants 1981, Waddington 1981, 1995, Dukas and Morse 2003, Suttle 2003, Makino and Sakai 2007, Ings and Chittka 2009), details concerning the mechanisms responsible for the decision rules are best figured out under con-
- 40 trolled laboratory conditions (Real 1981, Real et al. 1982, Chittka and Thomson 1997, Menzel 2001, Keasar et al. 2002, Chittka and Spaethe 2007).

The distribution of resources and predators are two of the main factors affecting choices of foraging animals.

- 45 Everything else being equal, animals tend to adopt the foraging strategies that maximise their intake rate and minimise their exposure to predators. When those strategies leading to higher intake rates are associated with higher intake rates, however, animals will only be selected to choose the rich-
- 50 risky strategy if the fitness returns of increasing intake rate are sufficient to compensate for the increased mortality rate
- 52 (reviewed by Gilliam and Fraser 1987, Lima and Dill 1990,

the rate of resource harvesting and minimizing predation risk has been studied in many taxa, but in the context of pollination responses to resource availability (Pleasants 1981, Real and Rathcke 1991, Makino and Sakai 2007) and predation risk (Dukas and Morse 2003, Suttle 2003, Dukas 2005) have been studied in isolation.

85 Pollinating insects have long been used as a model system to test predictions from optimal foraging theory because of the strong direct links between resource acquisition and fitness in these animals (Heinrich 1979, Pyke 1979, Seeley 1985, Dreisig 1995, Bosch and Kemp 2004). The effect 90 of predators on pollinator behaviour, on the other hand, was long neglected on the assumption that predation is too infrequent to affect the foraging strategy of pollinators (Pyke 1979, Miller and Gass 1985). Nevertheless, a number of studies over the last decade have shown that predators 95 can affect the foraging strategy of pollinators at the inflorescence, plant and patch levels (Dukas 2001, 2005, Dukas and Morse 2003, Muñoz and Arroyo 2004). These studies have shown that, through their non-consumptive effects on pollinator behaviour (Dukas and Morse 2003, Gonçalves-100 Souza et al. 2008, Ings and Chittka 2009), predators may have top-down effects on plant fitness and even affect the structure of the plant-pollinator community (Suttle 2003, Gonçalves-Souza et al. 2008). A recent model has shown that pollinator's choice of foraging strategy should be affected by 105 their susceptibility to predation, as well as the abundance

Brown and Kotler 2004). The tradeoff between maximizing

15

20

OIKO A 019910.indd 1

- 0 of resources and distribution of predators: less vulnerable pollinators should concentrate their foraging effort on rich, risky flowers and more vulnerable pollinators on poor, safe flowers (Jones 2010), while an earlier model demonstrated that social and solitary species should differ in their response
- 5 to predation risk (Clark and Dukas 1994). It is also known that the tendency of bumblebees, *Bombus occidentalis*, to flee from a model predator increases with the level of resources at the hive (Cartar 1991). Yet no study has investigated how pollinators trade off increasing resource acquisition and minimising predation risk.

The response of pollinators to the interplay between resource availability and predation risk may have ecological and evolutionary implications. Thus, when predators are relatively sedentary and flowers or inflorescences long-

- 15 lived (Morse 2007), resources will tend to accumulate in areas where predation risk is high if pollinators avoid them. Will the accumulation of resources tempt pollinators back into predator-rich patches? A similar question can be raised at the evolutionary time scale. Predators can show strong
- 20 preferences for some host plant species (Morse 2000, 2007, Schmalhofer 2001). Any plant species that recurrently experiences low reproductive success because it is used as hunting platform by ambush predators might be selected to increase reward production – at least if pollinators are willing to
- 25 increase their exposure to predation risk in order to increase the rate at which they acquire resources. The purpose of this paper is hence to study how pollinators trade off foraging efficiency for avoidance of predation risk. To tease apart the role of predators ambushing at flowers (hereafter referred to
- 30 as ambush predators) and resource availability from other floral traits, instead of comparing visit rates at flower species naturally differing in nectar production and the frequency with which they harbour ambush predators, we compared visit rates at flowers of a single species manipulated to differ
- 35 in their level of predation risk and resource availability. In particular, this experiment allows us to answer the following questions:
  - how do pollinators trade off between maximising intake rate and minimising predation risk?
  - 2) If pollinators avoid predator-rich areas, could inflorescences recover their attractiveness increasing their rate of nectar production?
- 45 The purpose of this study is not so much to learn how spatial heterogeneity in resource availability and predation risk affects plant–pollinator interactions in a particular community, as to understand the factors affecting the foraging strategies of pollinators. This insight can then be incorporated
- 50 into models and used to understand the ecological assemblage and evolutionary trajectories of pollination networks (Rodríguez-Gironés and Santamaría 2005, 2010).

# 55 Methods

40

#### Study site and species

We conducted our experiment in May 2007 in an abandoned crop field at 'la Raña' (39°41'51"N, 5°27'55"W) within 'Las Villuercas-Ibores' region in Extremadura, south-61 western Spain. The most common flowering plants at our study site were Chrysanthemum segetum (Asteraceae), Ornithopus compressus (Papilonaceae), Anthemis sp. (Asteraceae), Hedypnois cretica (Asteraceae), Leontodon taraxacoides (Aster-65 aceae), Echium plantagineum (Boraginaceae), Silene gallica (Caryophyllaceae) and Calendula arvensis (Asteraceae). We selected the field site because of the abundance of C. segetum inflorescences: they are commonly used by crab spiders as 70 hunting platform and are visited by a large number of nectar- and pollen-collecting insects, including Hymenoptera, Diptera, Coleoptera and Lepidoptera. In what follows, we consider only the interactions between honeybees Apis mellifera (Apidae), hoverflies Eristalis tenax (Syrphidae), hereafter 'hoverflies' and C. segetum, as they were the only ones 75 sufficiently common for statistical analysis.

We used adult females of two crab spider species, *Thomisus* onustus and Synema globosum (Thomisidae), as ambush predators in our experiment. Crab spiders are sit-and-wait ambush predators and use their enlarged powerful raptorial front legs to capture their prey (Morse 2007). *Thomisus onustus* and *S. globosum* prey mainly on bees and flies and were locally abundant in our field site.

#### **Experimental treatments**

We selected 40  $1 \times 1$  m<sup>2</sup> patches with high density of C. segetum for the experiment and mowed a 1.5 m wide strip of vegetation around each patch to decrease the number of crab spiders leaving the patch by bridging (Corcobado 90 et al. 2010). We grouped patches in ten blocks of four nearby patches each and allocated patches at random to the following treatments: rich-risky patches, poor-risky patches, richsafe patches and poor-safe patches, with one patch of each type per block. On 6 May we counted the number of pol-95 linators per patch four times, at 09:30, 11:30, 15:30 and 17:30. After the last count we counted and removed all the spiders we found in all the patches. We counted the number of C. segetum inflorescences per patch on 6, 13, 20 and 29 May and 1 June. 100

On 7 May we added seven *T. onustus* and three *S. globosum* females to each risky patch. From then on, we removed all spiders we found in safe patches during the experiment and we added crab spiders to risky patches whenever we found less than three individuals in a patch during the 105 observations.

For the nectar treatment we added 50  $\mu$ l of 15% (w/w) sucrose solution to 40 haphazardly selected C. segetum inflorescences (not harbouring spiders) in each rich patch twice per day, starting at 09:00 and 14:00. We used low concen-110 tration nectar because of the speed at which water evaporated from the exposed droplets. We chose the number of inflorescences to which we added sucrose solution and the amount of sucrose added per inflorescence so as to double nectar availability in rich and poor patches. Sucrose solu-115 tion was added with a repeater micropipette on the disc of inflorescence heads. Each day, twenty patches were observed between 10:00 and 13:00, and the remaining 20 after 15:00. In each half of the day we observed the 20 patches where nectar had been more recently added to minimize changes in nectar concentration. 121

2

- 0 We observed each patch for 15 min per day. We counted the number of pollinators present in the patch upon arrival of the observer and we recorded the number of insects visiting the patch and the number of inflorescences visited by each insect during the observation period. We
- 5 also recorded the number of spiders in the patch and how many of them were consuming prey at the start of the observations. Finally, we recorded the number of spider attacks and prey captures.

### 10 Strength of manipulations

The strength of the predation-risk manipulation was determined comparing the number of crab spiders present in risk and safe patches during the experiment with their natural den-

15 sity, which was estimated from the number of crab spiders we encountered on patches before the start of the experiment.

To determine whether spiders concentrated on patches with more resources, we used a paired t-test to compare the average number of spiders observed per patch between rich

20 and poor patches. We further compared the number of spiders that we had to add to or remove from rich and poor patches.

To assess the strength of the nectar enrichment manipulation, we assumed that nectar production rate by *C. segetum* 

- 25 is similar to that of *C. coronarium* in Greece, where each disk floret produces 0.01  $\mu$ l of 47% (w/w) nectar per day (Petanidou and Smets 1995). Because there are approximately 300 disk florets per inflorescence in *C. segetum* (Howarth and Williams 1972), nectar productivity must be about 3  $\mu$ l per
- inflorescence. Nectar productivity per patch was therefore estimated as the number of *C. segetum* inflorescences times 3 μl of 47% (w/w) per inflorescence. We used an ANOVA to test whether rich and poor patches differed in the number of inflorescences (averaged over the five counts) and hence in

35 the amount of resources they produced.

# Effect of predation risk and resource availability – patch level

- 40 For each patch, we averaged over five days of observations (15–19 May) the number of pollinators (honeybees and hoverflies) arriving to the patch during the observations, the number of inflorescences visited by each pollinator within the patch and the number of open inflorescences. These average
- 45 values were entered into mixed effects models to determine the effect of treatment and inflorescence abundance on pollinator behaviour. Each model included nectar presence (poor vs. rich patches), spider presence (safe vs risky patches) and their interaction as fixed factors, block (10 levels) as random
- 50 factor and number of inflorescences per patch as a covariate. The dependent variables were the number of pollinators (honeybees or hoverflies) and the average number of inflorescences visited per pollinator within the patch. To achieve homogeneity of variances, we square-root transformed the
- 55 number of inflorescences that each pollinator visited. Interactions between the number of inflorescences per patch and treatment (resource availability and/or predation risk) are not reported because they were never statistically significant and models including these interactions always lead to increases
- 60 in the AIC value greater than two units (Akaike 1973).

We first analysed the effect of predation risk and resource availability on the average number of visits that inflores-65 cences received. For each patch, we divided the total number of visits recorded during the observations (averaged over the five days of observations) by the number of inflorescences in the patch, thus obtaining the average number of visits per inflorescence. We analysed separately the data for honeybees 70 and hoverflies, using mixed effects models that included nectar presence (poor vs rich patches), spider presence (safe vs risky patches) and their interaction as fixed factors and block (10 levels) as random factor. We used Box-Cox transformations to achieve homogeneity of variances, with  $\lambda = 0.35$  for 75 the honeybees and  $\lambda = 0.5$  (equivalent to the square-root transformation) for the hoverflies.

We next focused on the response of pollinators to the presence of predators on the inflorescences they approached. To determine whether pollinator species and patch resource 80 availability affected the probability of landing on predator-harbouring inflorescences we used a generalized linear mixed effects model with a binomial distribution and identity link function. For each inflorescence visited in risky patches throughout the period of observations, the type of 85 inflorescence chosen (with or without predator) was used as dependent variable in the analysis. Resource availability (rich vs poor patches) and pollinator species (honeybees vs hoverflies) were the fixed factors, block (10 levels) was used as random factor and the proportion of inflorescences har-90 bouring spiders was used as covariate. If pollinators chose inflorescences at random, irrespective of the presence of spiders, there should be a linear relationship, with slope of one, between the proportion of inflorescences harbouring spiders and the probability of choosing a spider-harbouring inflo-95 rescence. If pollinators avoided inflorescences with spiders, the probability of choosing a spider-harbouring inflorescence might increase with the proportion of inflorescences harbouring spiders, but the slope of the relationship would be smaller than one. To test whether pollinators avoided spi-100 ders, we performed a Wald's Z test on the slope (Dobson and Barnett 2008), the null hypothesis being slope = 1.

### Effect of spider encounters on patch departure

Whenever a pollinator landed on a spider-harbouring inflorescence, we scored the response of the spider according to one of the following categories: indifference if the spider did not respond to the arrival of a pollinator, approach if the spider oriented and moved in the direction of the pollinator, 110 strike if the spider attempted to capture the pollinator with its forelegs and failed to contact the pollinator, struggle if the spider enclosed the pollinator with its forelegs but the pollinator managed to escape and kill if the spider managed to capture the pollinator. If the pollinator was not killed, 115 we further recorded whether the next inflorescence it visited was within the same patch. We used a generalised linear mixed effect model with binomial distribution to determine the factors affecting patch departure (stay vs leave patch). The model included spider response (still, approach, strike and struggle), resource availability (rich vs poor patches) and 121

61

0 pollinator species (honeybees vs hoverflies) as fixed factors and block (10 levels) as random factor.

#### Susceptibility to predation

- 5 We compared the susceptibility to predation of honeybees and hoverflies in two ways. First we used Fisher's exact test to compare 1) the proportion of honeybees and hoverflies that were attacked by spiders after landing on spider-harbouring inflorescences, and 2) from the attacked individuals, the pro-
- 10 portion that were actually captured. In this analysis, we only included those pollinators that landed on spider-harbouring inflorescences while we were observing the patch. In a second analysis, we estimated susceptibility to predation as the number of honeybees or hoverflies that were being consumed
- 15 by spiders when we arrived to a patch, normalised by the rate at which pollinators of the corresponding species visited the patch. (For each patch, we obtained a single value averaging over all observations.) This surrogate of susceptibility was then compared (honeybees vs hoverflies) with a Wilcoxon matched-pairs test.

Unless otherwise specified, all results are reported as average  $\pm$  SD, where averages refer to least squares means.

#### 25 **Results**

#### Strength of manipulations

Because *Chrysanthemum segetum* inflorescences in our patches received over 99% of honeybee and hoverfly visits to experimental patches during the observations, as a first approximation we can ignore other plant species when estimating resource availability per patch. There were no significant differences ( $F_{1,38} = 0.128$ , p = 0.721) between the number of *C. segetum* inflorescences at rich (175.4 ± 65.5) and

poor patches (182.3  $\pm$  55.5). At 3  $\mu$ l of nectar per inflorescence, the average patch offered 537  $\mu$ l of nectar. Given that the density of 15% and 47% sucrose solution is 1.06 and 1.22 g·cm<sup>-3</sup>, respectively, the 50  $\mu$ l of 15% sucrose solution

40 that we added to 40 inflorescences per rich patch correspond to 554  $\mu$ l of 47% nectar, implying that we essentially doubled the amount of nectar available at rich patches. On average, we added nectar to 22% of *C. segetum* inflorescences.

Before the manipulation, there were  $0.014 \pm 0.012$  crab spiders per inflorescence, with a range of 0 to 6 spiders per patch. During the observations, the number of spiders per inflorescence in risky patches was  $0.023 \pm 0.009$ , with 0 to 7 spiders per patch. Therefore the distribution of the num-

- ber of crab spiders per patch had similar ranges in risky and un-manipulated patches, although the average spider density in risky patches was 65% higher than the natural density in the area. We found and removed only 5 spiders from the safe patches throughout the experiment, so safe patches were essentially predator-free.
- The average number of spiders per patch was similar in poor  $(3.50 \pm 0.65)$  and rich  $(3.42 \pm 0.72)$  patches. The difference was not statistically significant (t<sub>9</sub> = 0.30, p = 0.77). During the observations, we added one spider per patch, with the exception of one rich and one poor patch to which we added four spiders. Likewise, of the five spiders

we removed from safe patches, two were in poor patches and 61 three in rich patches. Hence, rich and poor patches did not differ in their level of predation risk, or in the number of spiders they attracted.

Although resource availability and predation risk affected 65 the average number of pollinators visiting patches, they did not affect the range of visitors we encountered. Before the onset of the experimental manipulation, the number of visitors we encountered upon arrival to a patch was between 0 and 5 for honeybees and between 0 and 3 for hoverflies. 70 During the application of experimental treatments, the number of visitors we encountered upon arrival to patches was between 0 and 4 for both honeybees and hoverflies – so pollinator activity at experimental patches was well within natural levels. 75

# Effect of predation risk and resource availability – patch level

Honeybees preferred rich to poor patches ( $F_{1,9} = 68.79$ , 80 p < 0.001) and safe to risky patches ( $F_{1,9} = 42.73$ , p < 0.001), and their preference for rich patches was higher in safe than in risky patches (interaction term:  $F_{1,9} = 13.70$ , p = 0.001): the number of honeybees visiting rich-safe patches was more than double than the number of honeybees visiting poor-safe patches, while the number of honeybees visiting rich-risky patches was only 78% higher than the number visiting poor-risky patches (Fig. 1a). Patches with more inflorescences attracted more honeybees ( $F_{1,26} = 14.66$ , p < 0.001).

Honeybees visited more inflorescences per patch in safe than in risky patches ( $F_{1,9} = 19.62$ , p = 0.002), and in rich than in poor patches ( $F_{1,9} = 22.26$ , p = 0.001). The effect of the interaction between resource availability and predation risk on the number of inflorescences that honeybees visited per patch was not statistically significant ( $F_{1,9} = 1.79$ , p = 0.21; Fig. 1b). The number of inflorescences that honeybees visited before leaving a patch increased with the number of inflorescences in the patch ( $F_{1,26} = 9.79$ , p = 0.004).

Hoverflies preferred poor to rich patches ( $F_{1,9} = 6.94$ , 100 p = 0.01). Although the main effect of predation risk was not statistically significant (risk vs poor patches:  $F_{1.9} = 0.00$ , p = 0.96), the interaction between resource availability and predation risk had a statistically significant effect on the number of hoverflies visiting patches ( $F_{1,9} = 4.33$ , p = 0.047). 105 Although the numbers of hoverflies visiting rich-safe and poor-safe patches was similar, more hoverflies visited poorrisky than rich-risky patches. The number of hoverflies visiting patches was therefore smallest at rich-risky and highest at poor-risky patches (Fig. 2a). The number of hoverflies visit-110 ing patches increased with the number of inflorescences in the patch ( $F_{1,26} = 6.03$ , p = 0.02).

None of the factors studied had a clear effect on the number of inflorescences that hoverflies visited per patch. Hoverflies tended to visit more inflorescences in patches where 115 inflorescences were more abundant, but this trend was not statistically significant ( $F_{1,26} = 3.43$ , p = 0.075). Likewise, although the average number of inflorescences that hoverflies visited per patch was higher in rich than in poor patches, in safe than in risky patches, the effects of resource availability ( $F_{1,9} = 2.87$ , p = 0.12) and predation risk ( $F_{1,9} = 4.03$ , 121

OIKO\_A\_019910.indd 4





Figure 1. (a) Average rate (hour<sup>-1</sup>) at which honeybees, *Apis mellifera*, visited patches; (b) average number of inflorescences that individual honeybee visited before leaving the patch, and (c) rate (hour<sup>-1</sup>) at which the average inflorescence was visited by honeybees. Circles represent least-squared means ± SE for the four treatments (rich and poor, safe and risky patches; 10 replicas).

p = 0.076) did not reach statistical significance (Fig. 2b). The interaction between resource availability and predation risk had no discernable effects on the number of inflorescences that hoverflies visited per patch (F<sub>1,9</sub> = 0.05, p = 0.82).

#### Effect of predation risk and resource availability – inflorescence level

Both the number of honeybees visiting patches and the number of inflorescences that each honeybee visited per patch were greater in rich than poor patches, in safe than risky patches (Fig. 1a–b). As a result, there were statistically significant

Figure 2. (a) Average rate (hour<sup>-1</sup>) at which hoverflies, *Eristalis tenax*, visited patches; (b) average number of inflorescences that individual hoverflies visited before leaving the patch, and (c) rate (hour<sup>-1</sup>) at which the average inflorescence was visited by hoverflies. Circles represent least-squared means  $\pm$  SE for the four treatments (rich and poor, safe and risky patches; 10 replicas).

effects of resource availability ( $F_{1,9} = 54.19$ , p < 0.0001) and predation risk ( $F_{1,9} = 74.93$ , p < 0.0001), but not of their interaction ( $F_{1,9} = 3.28$ , p = 0.10), on the rate at which inflorescences were visited by honeybees. Note that inflorescences in risky-poor patches received less than half the number of honeybee visits per unit time than inflorescences in safe-poor patches, but inflorescences in risky-rich patches received as many honeybee visits as in safe-poor patches (Fig. 1c).

The pattern was different for hoverflies, as the number of inflorescences visited per hoverfly was lowest in the patches that received the greatest number of hoverfly visitors (Fig. 2a–b). This combination resulted in inflorescences receiving similar rates of hoverfly visits in all patch types (Fig. 2c). Neither resource availability ( $F_{1,9} = 0.02$ ,

5

60

105

110

115

- 0 p = 0.90), predation risk ( $F_{1,9} = 0.29$ , p = 0.60) or their interaction ( $F_{1,9} = 2.16$ , p = 0.18) had statistically significant effects on the rate at which inflorescences were visited by hoverflies.
- We now focus on those inflorescences where spiders were hunting. The probability that visitors to risky patches landed on spider-harbouring inflorescences was not significantly affected by patch type (rich or poor), pollinator species (honeybee or hoverfly) or their interaction (all  $p \ge 0.2$ ). However, due to the small number of visits to inflorescences
- 10 with spiders (61 out of 8081 pollinator visits) the test has relatively little power and the null hypotheses must be retained with caution. Despite the low proportion of visits to spider-harbouring inflorescences, the probability of landing on a spider-harbouring inflorescence increased with the
- 15 proportion of inflorescences within a patch which harboured spiders (p < 0.0001). The slope of this relationship, 0.33 (SE 0.06), was significantly smaller than one (W = -11.13, p < 0.001), indicating that both honeybees and hoverflies avoided spider-harbouring inflorescences. Avoidance of spi-
- 20 der-harbouring inflorescences becomes also apparent when we note that the proportion of inflorescences harbouring spiders in risky patches, 0.02, was greater than the proportion of visits to spider-harbouring inflorescences in risky patches, 0.0075. If honeybees and hoverflies were selecting inflores-
- 25 cences at random, the probability that they selected 61 or fewer spider-harbouring inflorescences out of 8081 landings would be  $5.6 \times \cdot 10^{-20}$  (binomial test).

#### Effect of spider encounters on patch departure

30 The tendency of honeybees and hoverflies to leave the patch following a non-lethal encounter with a spider increased as the response of the spider escalated from indifference through approach and strike to struggle (Fig. 3). The effect of spider response on the probability of leaving the patch was highly 35 significant (deviance = 21.10, DF = 3, p = 0.0001). All honeybees and hoverflies remained in the patch after encountering a spider that did not react to their landing, and left the patch after a struggle with a crab spider. On the other hand, neither visitor species (honeybee vs hoverfly) nor patch type 40 (rich vs. poor patches) had statistically significant effects on the probability of leaving the patch upon an encounter with a spider (species: deviance = 0.72, DF = 1, p = 0.40; resource availability: deviance = 0.85, DF = 1, p = 0.36).

45

50

#### Susceptibility to predation

Over 13 days of observations, we recorded 33 honeybees and 28 hoverflies landing on inflorescences harbouring crab spiders. Of these, 20 honeybees (60.61%) and 16 hoverflies (57.14%) were attacked by the spider. Spiders were therefore equally likely to attack honeybees and hoverflies (Fisher's exact test, two-tailed: p = 0.80). Of the 20 honeybees attacked, 9 (45%) were killed, while only 4 (25%) hoverflies were contured by spiders. Although the differ

55 hoverflies were captured by spiders. Although the difference in susceptibility was not statistically significant (Fisher's exact test, two-tailed: p = 0.30), the probability of detecting a significant difference with our sample size would be very low. Even if the observed capture frequencies (45% vs 24%) 60 represented the real susceptibility to predation of honeybees



Figure 3. Proportion of honeybees and hoverflies leaving the patch after a non-lethal encounter with a crab spider, plotted against the response of the spider. Sample sizes are indicated as (number of honeybees, number of hoverflies). Error bars represent SE, and are omitted when all individuals within a group showed the same response. 80

and hoverflies, over 80 honeybees and 80 hoverflies would have to be attacked before the probability of detecting a significant difference in success rate reached 50% (as calculated from 2000 Monte Carlo simulations for each sample size). 85 A more powerful test of susceptibility to predation is therefore obtained comparing the number of honeybees and hoverflies that spiders were eating when we arrived to the patches, normalised by the visit rate of the corresponding species. When we compared those prey that spiders were 90 consuming at the start of the observations, the proportion of visiting pollinators captured by spiders was higher for honeybees  $(0.15 \pm 0.12)$  than for hoverflies  $(0.04 \pm 0.03)$ , the difference being significant according to the Wilcoxon matched-pair test (Z = 3.88, p < 0.001, n = 20). How-95 ever, it is important to note that A. mellifera honeybees have a higher dry mass (29.5  $\pm$  1.1 mg) than *E. tenax* hoverflies (17.2  $\pm$  2.6 mg), which will possibly result in a longer spiders' handling time for bees compared to hoverflies (Brechbühl et al. 2010). A proper comparison of honeybee 100 and hoverfly susceptibility to predation therefore requires dividing the number of spiders consuming each prey type by the handling time of that prey type or, equivalently, dividing the number of spiders consuming honeybees by the ratio 'handling time for honeybees' / 'handling time for hoverflies'. 105 Using a ratio of 1.7 for the correction (which assumes a linear relationship between dry body mass and handling time, 29.5/17.2 = 1.7) we still find a highly significant difference between honeybee and hoverfly susceptibility (Z = 3.21, p = 0.001, n = 20). Indeed, the ratio of handling times 110 would have to be greater than 2.5 for the susceptibilities not to be significantly different at the 5% level (Fig. 4).

### Discussion

To the best of our knowledge, this is the first study of how pollinators trade off intake rate and predation risk. It allows us to answer the two questions we raised at the beginning of the study. 1) Honeybees and hoverflies responded to variations in predation risk and foraging success, albeit in completely 121



15 Figure 4. p-values for the comparison between honeybee and hover-fly susceptibility to predation (Wilcoxon matched pair test) versus the value by which the number of spiders consuming honeybees was divided to correct for the difference in handling times. Values greater than 1 imply that handling times are greater for honeybees than for hoverflies. The dashed line indicates the 0.05 significance level.

different ways. The most susceptible pollinators, honeybees, avoided risky patches, particularly if their profitability was low (Fig. 1a), while less susceptible hoverflies visited most often low-quality risky patches (Fig. 2a). 2) In the absence

- often low-quality risky patches (Fig. 2a). 2) In the absence of nectar addition, honeybee visit rates were lower for inflorescences in risky than in safe patches. Nevertheless, inflorescences in risky-rich patches received as many honeybee visits per unit time as inflorescences in safe-poor patches (Fig. 1c).
  Since our treatments did not affect the rate at which inflorescences.
- 50 Since our treatments did not affect the rate at which innorescences were visited by hoverflies (Fig. 2c), we can conclude that, while ambush predators make inflorescences less attractive to the pollinator ensemble, inflorescences can recover their attractiveness increasing nectar production rate.
- 35 Ambush predators can have positive and negative effects on the reproductive success of the plants they use as hunting platforms (Suttle 2003, Romero and Vasconcellos-Neto 2004, Gonçalves-Souza et al. 2008). When pollination limits the reproductive success of the plant, plants with ambush
- 40 predators are likely to experience a decrease in seed set (Higginson et al. 2010). Because ambush predators like crab spiders preferentially adopt certain flowers and inflorescences as hunting platforms (Morse 2007), it has been suggested that they can affect the population dynamics of their host plants
- 45 (Suttle 2003). The effect of crab spiders on their community will be all the most pronounced when, as we found, pollinators avoid not only spider-harbouring inflorescences, but their entire neighbourhood. At the evolutionary time scale, however, our results suggest a mechanism through which
- 50 flower species regularly associated with ambush predators could attract pollinators despite the increased predation risk: increasing reward production. Nectar availability in rich patches was roughly double than in natural patches. It is worth noting that, due to the way in which the experimental
- 55 manipulation was conducted, bees probably perceived rich patches less than 'twice as good' as poor patches. First of all, at the time when nectar was added it was relatively diluted (15% w/w) and therefore unattractive to bees. Although nectar would become more attractive as sugar concentration increased through water evaporation, the amount of sugar at

flowers also decreased through nectar consumption. Further-61 more, bees are risk-averse foragers: they prefer to visit patches where all flowers have similar amounts of nectar rather than patches with the same average amount of nectar per flower but higher inter-flower variance (Waddington et al. 1981, 65 Real 1981, Real et al. 1982). Because the proportion of C. segetum inflorescences to which we added nectar was on average 0.22, rich patches, as perceived by pollinators, were not as good as a hypothetical patch where each inflorescence 70 would have had twice as much nectar as in a normal patch. As a result, our nectar manipulation did not increase pollinator activity unrealistically in rich patches, but was sufficient to compensate for the presence of predators: inflorescences in safe-poor and risky-rich patches received similar amounts 75 of honeybee (Fig. 1c) or hoverfly (Fig. 2c) visit rates.

In is worth mentioning that the relationship between nectar availability and pollinator attractiveness in the presence of ambush predators can be quite complex. In our study, spiders did not respond to the nectar manipulation: we had 80 to add as many spiders to poor-risky and rich-risky patches, indicating that spiders were not more likely to leave poor than rich patches, and we had to remove as many spiders from poor-safe and rich-safe patches, indicating that spiders were not more likely to move to rich than poor patches. Nevertheless, other studies have shown that crab spiders show 85 a preference for more rewarding plant individuals (Heiling and Herberstein 2004) and species (Schmalhofer 2001). If flowers regularly used as hunting platforms by spiders increased nectar productivity to attract more pollinators, and the arrival of additional pollinators attracted more spiders 90 (Jones 2010, Abbott 2010), the system might enter a coevolutionary arms race with uncertain endpoint.

#### Different patterns at different spatial scales

At the flower level, honeybees and hoverflies showed similar responses: both species avoided spider-harbouring inflorescences. We found no significant differences between species in the probability of landing on spider-harbouring inflorescences while foraging in risky patches. Furthermore, the rates 100 observed for honeybees (33/4405 = 0.0075) and hoverflies (28/3676 = 0.0076) were so similar that any statistically significant difference that could be detected increasing sample size would be biologically irrelevant. Honeybees and hoverflies also reacted similarly to non-lethal encounters with crab 105 spiders. They tended to remain in the patch if the crab spider responded weakly to their presence, and to leave the patch after an attack (Fig. 3). The two species therefore exhibited strong anti-predator behaviour at the inflorescence level, and at this spatial scale the anti-predator response was not affected 110 by resource availability in the patch. Despite these similarities, when we analysed the foraging strategies of honeybees and hoverflies at the patch level we found striking differences in the number of individuals visiting patches (Fig. 1a, 2a). Between-species differences in patch-level response can have 115 their origin in mechanistic and functional differences.

#### Mechanisms: patch choice

Honeybees are central-place foragers. Numerous observations on marked bees indicate that workers concentrate their 121

7

- 0 foraging effort on a restricted area that they revisit trip after trip, even though each trip may include visits to inflorescences not belonging to the bee's core territory (Ribbands 1949, Free 1966). Moreover, individual honeybee and bumblebee workers have important learning and memory
- 5 capabilities that affect their foraging behaviour (Menzel 2001, Keasar et al. 2002, Giurfa 2007, Ings et al. 2009b) and they can learn to avoid specific flowers, flower species and even areas where they have been attacked (Abramson 1986, Dukas 2001, Ings and Chittka 2008, 2009). Recent
- laboratory experiments conducted under controlled conditions have shown that learning can play a key role in predator avoidance (Ings and Chittka 2008, 2009). Therefore, through the process of learning, rich and safe patches will be included in the foraging territories of more honeybees than
   poor and risky patches explaining why more bees visit rich

than poor patches, safe than risky patches (Fig. 1b).

Much less is known about the foraging ecology of hoverflies. Males acquire mating territories and are therefore residential (Wellington and Fitzpatrick 1981). If males

- 20 avoided spider-harbouring inflorescences and left patches upon attack by crab spiders, male territories would concentrate in safe patches. However, we rarely observed hoverflies behaving territorially. Most of our observations concerned foraging individuals that arrived to the patch and left it after
- 25 visiting a few inflorescences. Because the flight pattern of foraging *E. tenax* is characterised by a strong directionality (Gilbert 1983), non-territorial hoverflies are likely to wander through their environment without forming special attachments to any particular location. If this is the case, hoverflies
- 30 will have little or no information concerning the quality of the patches they approach. The number of hoverflies arriving to a patch must therefore be a function of the attractiveness of the patch, as assessed from whatever information hoverflies can obtain at a distance. Because resource availability
- 35 per flower and abundance of crab spiders cannot be detected at a distance, if hoverflies have no information concerning the patches they approach they must rely on other cues to select patches. Hoverflies showed a preference for patches with more *C. segetum* inflorescences – a trait that can be perceived from afar. As we discuss below, hoverflies may also
- have used the presence of other pollinators as a cue to assess the suitability of patches (Morse 1981).

#### Functional considerations: predator avoidance

45 From a functional point of view, honeybees may be avoiding risky patches because of their higher susceptibility to predation. Although there were no obvious differences in the ability of honeybees and hoverflies to detect and avoid 50 spider-harbouring inflorescences, honeybees were more vulnerable to predation than hoverflies once they landed on a spider-harbouring inflorescence. Schmalhofer (2001) suggested that the low representation of syrphids on the diet of *Misumenoides formosipes*, relative to honeybees, might be

- due to the clumsiness of honeybees (Fritz and Morse 1985) and the extreme agility and speed of syrphids (Barth 1991). Whatever the reason for the difference in susceptibility to predation between honeybees and hoverflies, susceptibility to predation is known to affect the predator-avoidance response of pollinators, in agreement with theoretical models (Jones
  - 8

2010). For example, Dukas and Morse (2003) reported that 61 while small and easily handled pollinators like B. ternarius and A. mellifera avoided crab spiders (Misumena vatia), larger pollinators like B. terricola and B. vagans did not show any anti-predatory response (see also Dukas and Morse 65 2005). Gonçalves-Souza et al. (2008) also found that not all pollinator species responded equally to the presence of an artificial crab spider sitting on Rubus rosifolius flowers: while hymenopterans tended to avoid flowers harbouring the artificial crab spider, lepidopterans did not show such a 70 response. These authors suggest that the absence of predatoravoidance mechanisms in butterflies could be due to their unpalatability. Indeed, we have observed in the field crab spiders eschewing butterflies after grabbing them with their forelegs. According to these and our study, pollinators suf-75 fering a low predation risk will show a weak anti-predatory response, while pollinators that are more vulnerable to pre-

#### Functional considerations: resource competition 80

dation will show stronger anti-predator behaviour.

Exploitation competition is known to play an important role in pollinator communities. To cite some examples, where Bombus appositus and B. flavifrons competed for the nectar produced by *Delphinium barbeyi* and *Aconitum colombianum*, 85 B. appositus concentrated its foraging effort on flowers of D. barbeyi and B. flavifrons on flowers of A. colombianum, but when one species was temporarily removed, the remaining bumblebee species increased visitation to the other flower species (Inouye 1978). In an experiment with marked bum-90 blebees, Thomson et al. (1987) found that when some bees were removed, remaining bees shifted their foraging activity towards the removal areas, thus increasing their foraging efficiency. Likewise, competition with honeybees forced B. occidentalis colonies to change their foraging strategy, allo-95 cating a greater fraction of their foragers from pollen to nectar collection (Thomson 2004). It is therefore possible that the high number of hoverflies visiting poor-risky patches reflects the fact that hoverflies are selecting to forage at those patches where honeybee activity is lower. Nevertheless, it is unclear 100 whether hoverflies and honeybees compete through the exploitation of resources or some form of territoriality: hoverflies, Melanostoma mellinum, foraged preferentially on flowers and patches where bumblebees, B. terricola and B. vagans, had been excluded, and avoided returning to flowers from which 105 they had been displaced by bumblebees (Morse 1981). It is therefore possible that, in our experiment, hoverflies were not choosing poor-risky patches to maximise their fitness. They may be simply excluded from rich-safe patches by honeybees. It is important to elucidate the mechanisms of resource com-110 petition between pollinator groups if we are to understand how pollination networks are structured.

Finally, other than competition for resources, honeybees and hoverflies may be reacting differently to our experimental treatments because they have different requirements. Honeybees must collect enough resources to sustain the growth of the colony during spring and summer, bringing enough pollen and nectar to feed non-foraging workers and developing larvae, and to keep the colony alive over the fall and winter (Seeley 1985). Hoverflies, on the other hand, require only resources for their own needs (including egg production, but 121

- 0 not larval growth). This life-history difference means that, to make ends meet, hoverflies can exploit resources where the average rate of gain is relatively low, while bees require much richer resources. Indeed, bumblebees rarely visit flowers where their average rate of gain is less than 0.02 W
- 5 (Heinrich 1975), while hoverflies accept resources with net energy intake rate of about 0.01 W (Gilbert 1983). If the productivity of *C. segetum* patches is just above the threshold for productive honeybee exploitation, a small increase in predation risk may suffice to tip the balance between exploi-
- 10 tation and neglect. By itself, however, it does not explain why hoverflies visited poor-risky patches at a higher rate than poor-safe and rich-safe patches. A combination of several factors (use of information, avoidance of competition and low energetic requirements) may be required to explain the

15 complex pattern of patch use by hoverflies.

# Conclusion

- 20 Our results show that hoverflies and honeybees responded differently to spatial variability in levels of resource availability and predation risk at the patch level, although both species strongly avoided spider-harbouring inflorescences and left patches following an attack. Although we have suggested
- 25 some mechanisms that can affect these behavioural differences, laboratory experiments, controlling for the previous exposure to food reward and predation risk of individual pollinators, are needed to elucidate the specific mechanisms by which pollinators respond to variability in predation risk
- 30 and food reward simultaneously. Whatever the mechanisms involved, ambush predators are likely to affect reproductive success of the flowers they use as hunting platforms (as suggested by Ings and Chittka 2009), but will also interfere with pollen flow in their immediate neighbourhood. It follows
- 35 that the reproductive success of a plant will not only depend on its phenotypic traits, but also on those of its neighbours. Therefore, the spatial scale at which predators and resource availability affect pollinator behaviour must be included in any ecological or evolutionary analysis of how predators affect
- 40 plant-pollinator interactions. In conclusion, our study highlights the importance of studying the foraging strategies of pollinators at different spatial scales in order to fully understand how plant-pollinator interactions are established.
- 45 Acknowledgements We thank Tina Calero, J. P. López, B. Llandres and I. López for technical support during the development of the experiment, L. Corcobado for help with the spider collection, P. Vargas for assistance with species identification, and L. Santamaría, J. Moya and O. Verdeny for statistical advice. We thank J. Bosch, J.
- 50 M Gómez, A. González-Megías, T. Ings, J. Moya-Laraño and M. Stang for or helpful comments to a previous draft of the manuscript. This work was supported by the Ministerio de Ciencia e Innovación/FEDER (projects CGL2007-63223/BOS and CLG2010-16795/BOS) and CSIC (project PIE 2006 3 OI 073 to MARG and studentship I3P-BPD2005 to ALL).
- 55

# References

 Abbott, K. R. 2010. Background evolution in camouflage systems: a predator–prey/pollinator–flower game. – J. Theor. Biol. 262: 662–678.

- Abramson, C. I. 1986. Aversive-conditioning in honeybees (*Apis 61 mellifera*). J. Comp. Psychol. 100: 108–116.
- Akaike, H. 1973. Information theory and an extension of the maximum likelihood principle. – Proc. 2nd Int. Symp. Information Theory, Budapest, pp. 267–281.
- Barth, F. G. 1991. Insects and flowers: the biology of a partnership. – Princeton Univ. Press.
- Bascompte, J. et al. 2003. The nested assembly of plant– animal mutualistic networks. – Proc. Natl Acad. Sci. USA 100: 9383–9387.
- Bosch, J. and Kemp, W. P. 2004. Effect of pre-wintering and wintering temperature regimes on weight loss, survival, and emergence time in the mason bee *Osmia cornuta* (Hymenoptera: Megachilidae). Apidologie 35: 469–479.
- Brechbühl, R. et al. 2010. Ineffective crypsis in a crab spider: a prey community perspective. Proc. R. Soc. B. 277: 739–746.
- Brown, J. S. and Kotler, B. P. 2004. Hazardous duty pay and the foraging cost of predation. – Ecol. Lett. 7: 999–1014.
- Cartar, R. V. 1991. Colony energy-requirements affect response to predation risk in foraging bumble bees. – Ethology 87: 90–96.
- Chittka, L. and Thomson, J. D. 1997. Sensori-motor learning and its relevance for task specialization in bumble bees. – Behav.
   Ecol. Sociobiol. 41: 385–398.
- Chittka, L. and Spaethe, J. 2007. Visual search and the importance of time in complex decision making by bees. – Arthropod–Plant Interactions 1: 37–44.
- Clark, C. W. and Dukas, R. 1994. Balancing foraging and antipredator demands - an advantage of sociality. – Am. Nat.144: 542–548.
- Corcobado, G. et al. 2010. Introducing the refined gravity hypothesis of extreme sexual size dimorphism. – BMC Evol. Biol. 10: 136.
- Dobson, A. and Barnett, A. 2008. An introduction to generalized linear models. – Chapman and Hall/CRC.
- Dreisig, H. 1995. Ideal free distributions of nectar foraging bumblebees. Oikos 72: 161–172.
- Dukas, R. 2001. Effects of perceived danger on flower choice by bees. Ecol. Lett. 4: 327–333.
- Dukas, R. 2005. Bumble bee predators reduce pollinator density and plant fitness. – Ecology 86: 1401–1406.
- Dukas, R. and Morse, D. H. 2003. Crab spiders affect flower visitation by bees. Oikos 101: 157–163.
- Dukas, R. and Morse, D. H. 2005. Crab spiders show mixed effects on flower-visiting bees and no effect on plant fitness components. – Ecoscience 12: 244–247.
- Free, J. B. 1966. The foraging behaviour of bees and its effect on the isolation and speciation of plants. – In: Hawkes, J. G. (ed.), Reproductive biology and taxonomy of vascular plants. Pergamon Press, pp. 76–92.
- Fritz, R. S. and Morse, D. H. 1985. Reproductive success and foraging of the crab spider *Misumena vatia*. Oecologia 65: 194–200.
- Gilbert, F. S. 1983. The foraging ecology of hoverflies (Diptera, Syrphidae) - circular movements on composite flowers.
  Behav. Ecol. Sociobiol. 13: 253–257.
- Gilliam, J. F. and Fraser, D. F. 1987. Habitat selection under predation hazard - test of a model with foraging minnows. – Ecology 68: 1856–1862.
- Giurfa, M. 2007. Behavioral and neural analysis of associative learning in the honeybee: a taste from the magic well.
  J. Comp. Physiol. A. Neuroethol. Sens. Neural. Behav. 115 Physiol. 193: 801–824.
- Gonçalves-Souza, T. et al. 2008. Trait-mediated effects on flowers: artificial spiders deceive pollinators and decrease plant fitness. – Ecology 89: 2407–2413.
- Heiling, A. M. and Herberstein, M. E. 2004. Floral quality signals lure pollinators and their predators. – Ann. Zool. Fenn. 41: 421–428.

9

95

- 0 Heinrich, B. 1975. The role of energetics in bumblebee-flower interrelationships. – In: Gilbert, L. E. and Raven, P. H. (eds), Coevolution of animals and plants. Univ. of Texas Press.
  - Heinrich, B. 1979. Bumblebee economics. Harvard Univ. Press.
- 5 Higginson, A. D. et al. 2010. The impact of flower-dwelling predators on host plant reproductive success. – Oecologia 164: 411–421.
  - Howarth, S. E. and Williams, J. T. 1972. Chrysanthemum segetum L. J. Ecol. 60: 573–584.
- Ings, T. C. and Chittka, L. 2008. Speed-accuracy tradeoffs and false alarms in bee responses to cryptic predators. – Curr. Biol. 18: 1520–1524.
  - Ings, T. C. and Chittka, L. 2009. Predator crypsis enhances behaviourally mediated indirect effects on plants by altering bumblebee foraging preferences. – Proc. R. Soc. B 276: 2031–2036.
- 15 Ings, T. C. et al. 2009a. Ecological networks beyond food webs. – J. Anim. Ecol. 78: 253–269.
  - Ings, T. C. et al. 2009b. A population comparison of the strength and persistence of innate colour preference and learning speed in the bumblebee *Bombus terrestris*. – Behav. Ecol. Sociobiol. 63: 1207–1218.
- Inouye, D. W. 1978. Resource partitioning in bumblebees experimental studies of foraging behavior. – Ecology 59: 672–678.
  - Jones, E. I. 2010. Optimal foraging when predation risk increases with patch resources: an analysis of pollinators and ambush predators. – Oikos 119: 835–840.
  - Jordano, P. 1987. Patterns of mutualistic interactions in pollination and seed dispersal - connectance, dependence asymmetries and coevolution. – Am. Nat. 129: 657–677.
    - Keasar, T. et al. 2002. Bees in two-armed bandit situations: foraging choices and possible decision mechanisms. – Behav. Ecol. 13: 757–765.
    - Lima, S. L. and Dill, L. M. 1990. Behavioral decisions made under the risk of predation - a review and prospectus. – Can. J. Zool. Rev. Can. Zool. 68: 619–640.
- Makino, T. T. and Sakai, S. 2007. Experience changes pollinator responses to floral display size: from size-based to reward-based foraging. – Funct. Ecol. 21: 854–863.
  - Menzel, R. 2001. Behavioral and neural mechanisms of learning and memory as determinants of flower constancy. – In: Chittka, L. and Thomson, J. D. (eds), Cognitive ecology of pollination: animal behaviour and floral evolution. Univ. Press, pp. 21–40.
  - Miller, R. S. and Gass, C. L. 1985. Survivorship in hummingbirds - is predation important. – Auk 102: 175–178.
    - Morse, D. H. 1981. Interactions among syrphid flies and bumblebees on flowers. – Ecology 62: 81–88.
- 45 Morse, D. H. 2000. The role of experience in determining patchuse by adult crab spiders. – Behaviour 137: 265–278.
  - Morse, D. H. 2007. Predator upon a flower: life history and fitness in a crab spider. – Harvard Univ. Press.
  - Muñoz, A. A. and Arroyo, M. T. K. 2004. Negative impacts of a vertebrate predator on insect pollinator visitation and seed output in *Chuquiraga oppositifolia*, a high Andean shrub.
  - Oecologia 138: 66–73.

- Petanidou, T. and Smets, E. 1995. The potential of marginal lands 61 for bees and apiculture - nectar secretion in Mediterranean shrublands. – Apidologie 26: 39–52.
- Pleasants, J. M. 1981. Bumblebee response to variation in nectar availability. Ecology 62: 1648–1661.
- Pyke, G. H. 1979. Optimal foraging in bumblebees: rule of movements between flowers within inflorescences. – Anim. Behav. 27: 1167–1181.
- Real, L. A. 1981. Uncertainty and pollinator–plant interactions the foraging behavior of bees and wasps on artificial flowers. – Ecology 62: 20–26.
- Real, L. A. and Rathcke, B. J. 1991. Individual variation in nectar production and its effect on fitness in *Kalmia latifolia*. – Ecology 72: 149–155.
- Real, L. et al. 1982. On the tradeoff between the mean and the variance in foraging effect of spatial-distribution and color preference. Ecology 63: 1617–1623.
- Ribbands, C. R. 1949. The foraging method of individual honeybees. – J. Anim. Ecol. 18: 47–66.
- Rodríguez-Gironés, M. A. and Santamaría, L. 2005. Resource partitioning among flower visitors and evolution of nectar concealment in multi-species communities. Proc. R. Soc. B 272: 187–192.
- Rodríguez-Gironés, M. A. and Santamaría, L. 2010. How foraging behaviour and resource partitioning can drive the evolution of flowers and the structure of pollination networks. – Open Ecol. J. 3: 1–11.
- Romero, G. Q. and Vasconcellos-Neto, J. 2004. Beneficial effects 85 of flower-dwelling predators on their host plant. – Ecology 85: 446–457.
- Schmalhofer, V. R. 2001. Tritrophic interactions in a pollination system: impacts of species composition and size of flower patches on the hunting success of a flower-dwelling spider. – Oecologia 129: 292–303.
- Seeley, T. D. 1985. Honey bee ecology: a study of adaptation in social life. – Princeton Univ. Press.
- Suttle, K. B. 2003. Pollinators as mediators of top–down effects on plants. Ecol. Lett. 6: 688–694.
- Thomson, D. 2004. Competitive interactions between the invasive European honey bee and native bumble bees. – Ecology 85: 458–470.
- Thomson, J. D. et al. 1987. Response of traplining bumble bees to competition experiments - shifts in feeding location and efficiency. – Oecologia 71: 295–300.
- Waddington, K. D. 1981. Factors influencing pollen flow in bumblebee-pollinated *Delphinium virescens*. – Oikos 37: 153–159.
- Waddington, K. D. 1995. Bumblebees do not respond to variance in nectar concentration. – Ethology 101: 33–38.
- Waddington, K. D. et al. 1981. Floral preferences of bumblebees (*Bombus edwardsii*) in relation to intermittent versus continuous rewards. – Anim. Behav. 29: 779–784.
- Waser, N. M. and Price, M. V. 1998. What plant ecologists can learn from zoology. – Perspect. Plant Ecol. Evol. Syst.1: 137–150.
- Wellington, W. G. and Fitzpatrick, S. M. 1981. Territoriality in the drone fly, *Eristalis tenax* (Diptera, Syrphidae). – Can. Entomol. 113: 695–704.

115

121

65

70

75

90

55

50

25

30

40

